

Impact of sensor failure on the observability of flow dynamics at the Biosphere 2 LEO hillslopes

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Abstract

1 The Biosphere 2 Landscape Evolution Observatory (LEO) has been developed to investigate hydrological, chemical,
2 biological, and geological processes in a large-scale, controlled infrastructure. The experimental hillslopes at LEO
3 are instrumented with a large number of different sensors that allow detailed monitoring of local and global dynamics
4 and changes in the hydrological state and structure of the landscapes. Sensor failure, i.e., a progressive reduction in
5 the number of active or working sensors, in such an evolving system can have a dramatic impact on observability of
6 flow dynamics and estimation of the model parameters that characterize the soil properties. In this study we assess the
7 retrieval of the spatial distributions of soil water content and saturated hydraulic conductivity under different scenarios
8 of heterogeneity (different values of correlation length of the random field describing the hydraulic conductivity)
9 and a variable number of active sensors. To avoid the influence of model structural errors and measurement bias,
10 the analysis is based on a synthetic representation of the first hydrological experiment at LEO simulated with the
11 physically-based hydrological model CATHY. We assume that the true hydraulic conductivity is a particular random
12 realization of a stochastic field with lognormal distribution and exponential correlation length. During the true run,
13 we collect volumetric water content measurements at an hourly interval. Perturbed observations are then used to
14 estimate the total water storage via linear interpolation and to retrieve the conductivity field via the ensemble Kalman
15 filter technique. The results show that when less than 100 out of 496 total sensors are active, the reconstruction of
16 volumetric water content may introduce large errors in the estimation of total water storage. In contrast, retrieval of
17 the saturated hydraulic conductivity distribution allows the CATHY model to reproduce the integrated hydrological
18 response of LEO for all sensor configurations investigated.

Keywords: Sensor failure analysis; Data assimilation; Landscape Evolutionary Observatory; Soil moisture;

19 **1. Introduction**

20 Determination of the number and location of sensors needed to monitor a real-world hydrological process is a
21 classical problem in experimental design, where the best compromise between maximum amount of information and
22 minimum number of sensors is sought (e.g., [1, 2]). In this framework, an aspect that is rarely taken into consideration
23 is that sensors may fail during long-term experiments, thereby putting at risk the observability of the system since it
24 may not always be possible to replace broken sensors. The lifetime of sensors is thus a crucial unknown in experiments
25 of long duration, and it becomes important to be able to predict how the information obtained from the active sensors
26 changes over time as the sensor network deteriorates.

27 This is the premise for the present study, which is based on the setup of the Landscape Evolution Observatory
28 (LEO) of the Biosphere 2 facility near Tucson, Arizona. The three synthetic, controlled hillslopes at LEO were
29 constructed with the aim of improving our predictive understanding of the coupled physical, chemical, biological,
30 and geological processes at Earth's surface in changing climates [3]. Each hillslope is 30 m long and 11.15 m wide
31 and has an average slope of 10 degrees. The 1 m deep soil consists of basaltic tephra, ground to homogeneous
32 loamy sand texture. For the first years of LEO operation, vegetation is not present and the research is focused on the
33 characterization of the hydrological response of the hillslopes in terms of water transit times, generation of seepage
34 and overland flow, internal dynamics of soil moisture, and evaporation. The second part of the experiment envisages
35 the presence of plants growing on the hillslopes and aims to monitor the oxygen and carbon cycles inside LEO, as
36 well as the impact of vegetation on the spatial distribution of soil water content and on changes in the soil hydraulic
37 properties [4, 5].

38 To monitor these processes, each hillslope is equipped with a dense network of soil sensors (496 locations) that
39 measures volumetric water content (496 sensors), soil water potential, and soil temperature. These local observations
40 of the internal state of the soil are combined with measurements of the global system response, such as the total

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41 weight of the infrastructure (and thus the water storage), the rate of irrigation/evaporation, and the water outflow at
42 the seepage face. Finally, geochemical analysis of irrigation water, soil water, and seepage outflow are available to
43 monitor solute transport processes along the hillslopes.

44 As sensors fail, the number of active sensors, m , will decrease in time. For example, assuming that the time of
45 failure of a sensor, t_f , follows a Gamma distribution with shape parameter α and rate parameter β , the probability that
46 a sensor is working at time t is

$$47 \quad P_w(t) = P(t_f > t) = 1 - \int_0^t g(\tau; \alpha, \beta) d\tau \quad (1)$$

48 where $g(\cdot; \alpha, \beta)$ is the probability density function (pdf) of the Gamma distribution. With the further assumption that
49 the times of failure of the sensors are independent and identically distributed random variables, the number of active
50 sensors at time t has a binomial distribution with parameters $p = P_w(t)$ and $n = 496$, and the expected value of active
51 sensors at time t is $E[m] = np$. Figure 1 shows the probability of the lifetime of a sensor, $P_w(t)$, for two possible
52 combinations of parameters α and β (the expected value of the failure time in this example is $E[t_f] = \alpha/\beta = 10$ years).

53 In this study we assess the impact of the number of active sensors on the observability of the LEO hillslopes.
54 The physically-based hydrological model CATHY [6] is employed to numerically simulate the water dynamics on
55 the LEO landscapes. CATHY couples a finite element solver of the Richards equation for subsurface flow developed
56 by Paniconi and Putti [7] with a surface routing scheme developed by Orlandini and Rosso [8]. Surface flow occurs
57 along a conceptual channel network derived from the digital elevation model (DEM) of the landscape [9], and the
58 coupling between the surface and subsurface modules is resolved via a boundary condition-based partitioning of the
59 atmospheric inputs into soil infiltration and land surface ponding. To account for heterogeneities in the LEO soil [10],
60 we represent the saturated hydraulic conductivity as a three-dimensional random field with a lognormal probability
61 distribution and an anisotropic exponential covariance function.

62 We use two different approaches to quantitatively assess the information associated with the network of active
63 sensors of volumetric water content. In the first approach we are interested in knowing if LEO's sensor network
64 allows us to accurately retrieve the spatial and temporal distribution of the water content in the entire landscape. To
65 assess the accuracy of the retrieval, we compare the integral of the computed water content over the entire domain with
66 the measured variation of water storage in the landscape. In the second approach we assess the sensor network's ability

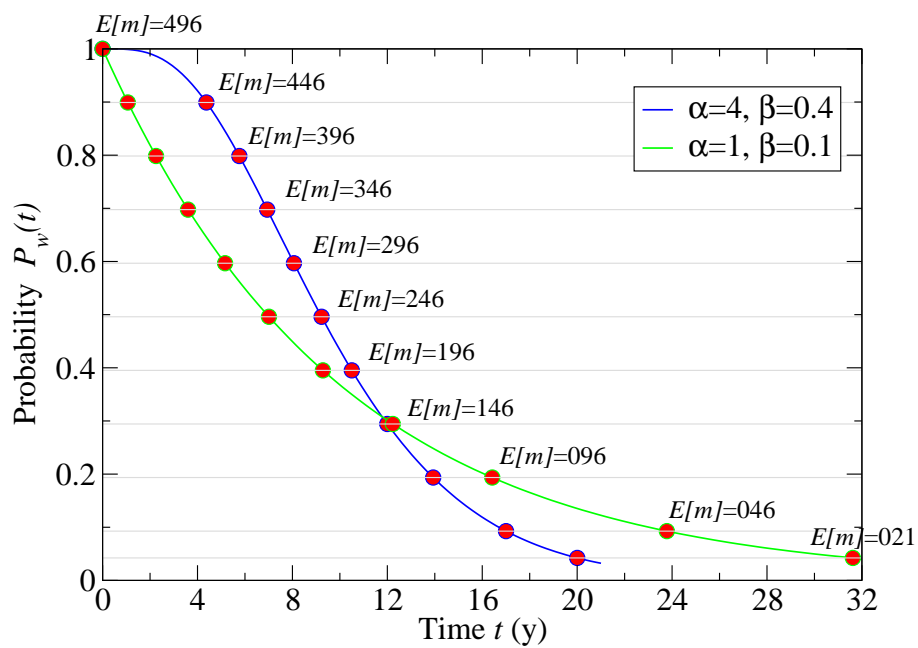


Figure 1: The probability of a sensor being active at time t for two distributions of failure time. The two red circles along each horizontal line give, for each distribution, the time at which the number of active sensors is expected to have dropped to the indicated value of $E[m]$. For instance, the expected value of active sensors is 46 ($E[m] = P_w n = 0.093 \times 496$) after 17 y for the blue distribution and after 23.8 y for the green distribution.

67 to allow retrieval of the saturated hydraulic conductivity of the soil, a critical parameter for numerical modeling of
68 the future hydrological experiments at LEO. To account for parameter and measurement uncertainties, the ensemble
69 Kalman filter (EnKF) [11–14] is used to compute the posterior probability distribution of the saturated hydraulic
70 conductivity. EnKF performs a Gaussian approximation of sequential Bayesian inversion, thereby extending the
71 Kalman filter to nonlinear models. The evolution in time of the state pdf is simulated using a Monte Carlo (MC)
72 technique. The ensemble of model solutions is associated with random realizations of the unknown parameters. These
73 MC realizations are then used in the update step to compute the covariance matrices required in the Kalman filter. Due
74 to its straightforward implementation and its computational efficiency [15], EnKF is largely employed in engineering
75 applications for measurement assimilation in real time. Moreover, since EnKF seeks a probability distribution of the
76 parameters, this approach reduces the issues associated with non-uniqueness of the solution that typically occurs in
77 inverse problems (e.g, [16]).

78 One of the major drawbacks of the EnKF technique is the so-called ensemble inbreeding (i.e., the strong reduction
79 of the ensemble variance after few updates). For this reason, Drecourt et al. [17] and De Lannoy et al. [18] suggest
80 that it is important to ensure that the ensemble spread is large enough at the assimilation time. Recent enhancements
81 to the EnKF technique for estimation of two-dimensional stochastic parameters include introduction of a damping
82 parameter [19] to reduce ensemble inbreeding, and covariance localization to clean the ensemble covariance matrices
83 of spurious terms [20, 21]. Sun et al. [22, 23] combine EnKF with grid-based localization and Gaussian mixture
84 model clustering techniques to estimate a multimodal parameter distribution. Panzeri et al. [24] couple EnKF with
85 the ensemble moment equation of the transient groundwater flow equation to circumvent the MC simulation. Alzraiee
86 et al. [25] compare centralized and decentralized fusion to invert the measurements generated with different pumping
87 tests. Amongst applications of EnKF for estimating the spatial distribution of parameters in three-dimensional hydro-
88 logical models, Chen and Zhang [26] showed that EnKF provides a satisfactory estimation of the three-dimensional
89 hydraulic conductivity field assimilating measurements of pressure head in a synthetic example of saturated flow.

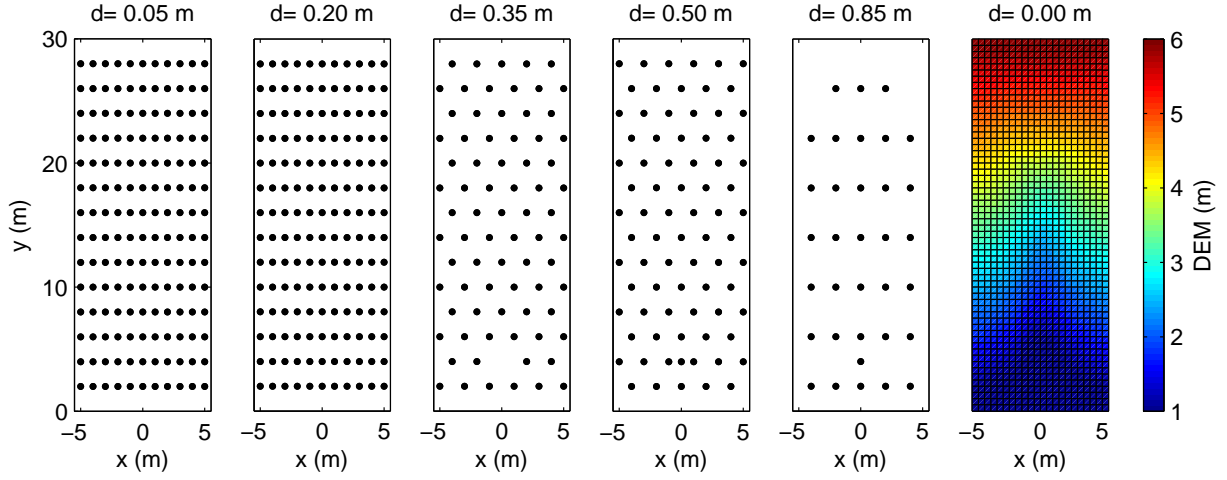


Figure 2: Digital elevation model of the surface of the hillslope domain Ω and the locations $\vec{x}_1, \dots, \vec{x}_m$ of the $m=496$ sensors of water content, installed at five different depths d .

90 2. Problem representation

91 We represent the hillslope (the three LEO hillslopes are identical) as a three-dimensional domain Ω with the DEM
 92 depicted in Figure 2 and a 1 m deep soil. The bottom of the hillslope, the two side boundaries (the edges along the
 93 y axis in Figure 2), and the upslope boundary are impermeable, while the downslope boundary (at $y=0$ m, hereafter
 94 denoted by Γ) is the outflow face, and is modeled as a seepage face boundary condition. Let $\theta(t, \vec{x})$ be the soil water
 95 content $[-]$ at a time $t [T]$ at a point $\vec{x} = (x, y, z) \in \Omega$. Given a spatial distribution of θ at a reference time $t_0=0$ (initial
 96 condition), rainfall and evaporation boundary conditions are imposed at the surface, and θ responds according to this
 97 forcing term and to the soil hydraulic properties.

98 The dense sensor network allows the system to be monitored every 15 minutes from the reference time t_0 (times t_i
 99 with $t_i - t_{i-1}=15$ min). To test the procedure with a known distribution of water content, in this study we consider the
 100 following synthetic measurements generated with the numerical model CATHY:

- 101 1. The outgoing water volume V_Γ at the seepage face $\Gamma [L^3]$,

$$102 \quad V_\Gamma(t_i) = \int_{t_{i-1}}^{t_i} \int_\Gamma K_S(\vec{x}) \cdot \nabla \psi(t, \vec{x}) \cdot \vec{n}_\Gamma(\vec{x}) d\vec{x} dt \quad (2)$$

103 where ψ is the pressure head, K_s is the saturated hydraulic conductivity tensor $[L/T]$, and \vec{n}_Γ is the outward
 104 normal vector at Γ .

105 2. The water storage V_Ω [L^3], computed as the time variation of the total weight P_Ω of the LEO infrastructure
 106 measured by 10 load cells:

$$107 \quad V_\Omega(t_i) = \int_\Omega \theta(t_i, \vec{x}) d\vec{x} = V_\Omega(t_0) + \gamma(P_\Omega(t_i) - P_\Omega(t_0)) \quad (3)$$

108 where γ is the specific weight of water.

109 3. The measurements of the soil water content $\mathbf{y}_i^o = \{\theta^o(t_i, \vec{x}_j^o)\}_{j=1}^{496}$ at 496 sensors (5TM Decagon probes). The
 110 sensor locations, $\vec{x}_1^o, \dots, \vec{x}_{496}^o$, are distributed at 5 depths within the domain (0.05, 0.20, 0.35, 0.50, and 0.85 m)
 111 as shown in Figure 2.

112 These measurements are subject to error. Measurement error on V_Γ and V_Ω is modeled as a multiplicative random
 113 noise ϵ with a lognormal distribution, $\xi \sim \text{LogN}$, unitary expected value, $E[\xi] = 1$, and coefficients of variation cv_{V_Γ}
 114 and cv_{V_Ω} for V_Γ and V_Ω , respectively. For example:

$$V_\Omega^o(t_i) = V_\Omega(t_i)\xi(t_i), \quad (4)$$

115 where V_Ω^o is the observed storage, V_Ω is the real storage and

$$\log(\xi(t_i)) \sim N(-0.5\log(1 + cv_{V_\Omega}^2), \log(1 + cv_{V_\Omega}^2)). \quad (5)$$

116 This approach, suggested by Camporese et al. [27], guarantees the positivity of the perturbed measurements.

117 The volumetric water content measurement error is modeled as an additive Gaussian process with mean 0 and
 118 variance σ_θ^2 , according to the calibration of the sensors:

$$\theta^o(t_i, \vec{x}_j^o) = \theta(t_i, \vec{x}_j^o) + \epsilon_j(t_i), \quad (6)$$

119 with

$$\epsilon_j(t_i) \sim N(0, \sigma_\theta^2). \quad (7)$$

120 Note that the perturbed measurements lower than the residual moisture content, θ_r [L^3/L^3], or higher than the saturated
 121 moisture content, θ_s [L^3/L^3], are corrected to the range limits.

122 Considering only the measurements of volumetric water content, the internal observability of the system depends
 123 on the number of active sensors m and on their spatial distribution σ_m , where σ_m is a possible selection of m sensors

124 among the initial 496 sensors. The number of possible combinations σ_m is the binomial coefficient of Newton $\binom{496}{m}$.

125 This is a huge number (e.g., $\binom{496}{446} \approx 10^{69}$, $\binom{496}{21} \approx 5 \cdot 10^{36}$). To take into account the influence of the spatial
 126 distribution of the m sensors in light of the impossibility of considering all possible combinations, we conduct the
 127 sensor failure analysis over a fixed number r of random distributions $\sigma_{m,1}, \dots, \sigma_{m,r}$.

128 We propose two methodologies to assess the critical number of active sensors, m^* , that might compromise the
 129 observability of the LEO system. In the first approach we employ the measurements of volumetric water content
 130 to evaluate the spatial distribution of θ over the entire domain, thus providing an estimate of the total storage of
 131 the system. In the second approach, the measurements are employed for the calibration of the saturated hydraulic
 132 conductivity in a numerical model of LEO.

133 3. Estimation of volumetric water content

134 Local measurements of water content allow us to infer the water content $\tilde{\theta}(t_i, \vec{x})$ at any point $\vec{x} \in \Omega$ by interpolation.
 135 To ensure monotonicity between the measured values, here we compute $\tilde{\theta}(t_i, \vec{x})$ with linear interpolation

$$136 \quad \tilde{\theta}(t_i, \vec{x}) = \sum_{j=1}^m \theta^p(t_i, \vec{x}_{\sigma_{m,r}(j)}) \phi_{\sigma_{m,r}(j)}(\vec{x}) \quad (8)$$

137 where $\phi_{\sigma_{m,r}(j)}(\vec{x})$ are piecewise linear interpolation functions such that $\phi_{\sigma_{m,r}(j)}(\vec{x}_l) = \delta_{\sigma_{m,r}(j),l}$. The functions $\phi_{\sigma_{m,r}(j)}$ are
 138 defined on the Delaunay triangularization associated with the location of the active sensors. The water content outside
 139 the convex hull delimited by the sensors is approximated with linear extrapolation and eventually corrected to the
 140 physical limits $[\theta_r, \theta_s]$. The subscript i in (8) refers to the measurement time, and $\sigma_{m,r}(j)$ indicates the j -th sensor
 141 among the m active sensors in the combination $\sigma_{m,r}$.

142 A simple control to check the reliability of the estimated water distribution at a time t_i consists in comparing the
 143 estimated water volume, $\tilde{V}_\Omega(t_i) = \int_\Omega \tilde{\theta}(t_i, \vec{x}) d\vec{x}$, with that measured by the load cells $V_\Omega(t_i)$. If the estimated water
 144 volume $\tilde{V}_\Omega(t_i)$ falls outside the 90% confidence interval of the measure, then the distribution of sensors $\sigma_{m,r}$ is not
 145 trustworthy for the estimation of water content in Ω .

146 3.1. Numerical model

147 The CATHY (CATchment HYdrology) model [6] has been developed for simulation of water dynamics in catch-
 148 ments and hillslopes. The model solves the three-dimensional Richards equation describing the dynamics of pressure
 149 head ψ in variably saturated porous media:

$$150 \quad S_w(\psi)S_s \frac{\partial \psi}{\partial t} + \theta_s \frac{\partial S_w(\psi)}{\partial t} = \nabla \cdot [K_s K_r(S_w(\psi)) (\nabla \psi + \eta_z)] + q \quad (9)$$

151 where S_s is the aquifer specific storage coefficient [L^{-1}], $S_w = \theta/\theta_s$ is the saturation [L^3/L^3], K_r is the relative
 152 hydraulic conductivity [-], ∇ is the gradient [$1/L$], and $\eta_z = (0, 0, 1)^T$, with z the vertical coordinate directed upward
 153 [L]. The term q represents source/sink fluxes [L^3/L^3T] internal to the domain Ω or forcing terms that control the
 154 fluxes along the domain boundary $\partial\Omega$.

155 Retention curves establish a one-to-one relationship between the pressure head ψ and the water content θ (hystere-
 156 sis is not considered in this work). The characteristics of the porous media, such as the pore size distribution n_p [-]
 157 and the pore entry suction $1/\alpha$ [L], determine the local shape of these curves. Here the van Genuchten curves [28] are
 158 employed to model the water content and the relative hydraulic conductivity K_r as a function of ψ :

$$159 \quad \theta(\psi) = \begin{cases} \theta_r + \frac{\theta_s - \theta_r}{(1 + (\alpha|\psi|)^{n_p})^{m_p}} & \text{if } \psi < 0, \\ 1 & \text{if } \psi \geq 0; \end{cases} \quad (10)$$

$$160 \quad K_r(\psi) = \begin{cases} \sqrt{1 + (\alpha|\psi|)^{n_p}} (1 - (1 - \frac{\theta(\psi) - \theta_r}{\theta_s - \theta_r})^{m_p})^2 & \text{if } \psi < 0, \\ 1 & \text{if } \psi \geq 0; \end{cases} \quad (11)$$

162 where $m_p = 1 - 1/n_p$. If saturation or infiltration excess runoff occurs, CATHY couples the Richards equation solver
 163 with a one-dimensional diffusion wave approximation of the Saint-Venant equation for overland flow routing. The
 164 mass balance at the surface/subsurface interface is enforced by a boundary condition switching algorithm described
 165 in [6].

166 CATHY can simulate the main outputs measured at LEO, such as the seepage face outflow V_Γ , the water storage
 167 V_Ω , and, in case of overland flow, the hydrograph at the outlet of the landscape.

168 3.2. State space model

169 After discretizing the domain Ω with n nodes and e tetrahedral elements, the numerical solution of equation (9) is
170 obtained by way of the finite element method (FEM) with piecewise linear basis functions [6]. The time-integration is
171 performed with the backward Euler method combined with Picard or Newton iterations. To simplify the notation, in
172 the following the CATHY state-space model is described considering only the subsurface processes. A more general
173 notation should include in the state vector also the surface discharge and the ponding water volumes, as described
174 by Pasetto et al. [29]. We indicate with

175
$$\mathbf{x}_i = \{\psi_k(t_i)\}_{k=1}^n \quad (12)$$

176 the numerical solution, i.e., the vector of pressure heads ψ_k at the location of the grid nodes at time t_i . The state
177 vector of the model, \mathbf{x}_i , can be formally expressed as a nonlinear function, \mathcal{F} , of the pressure head \mathbf{x}_{i-1} at the previous
178 observation time, of the forcing term q_{i-1} , and of the vector $\boldsymbol{\gamma}$ representing the soil properties (in our case the saturated
179 hydraulic conductivity K_S):

180
$$\mathbf{x}_i = \mathcal{F}(\mathbf{x}_{i-1}, q_{i-1}, \boldsymbol{\gamma}). \quad (13)$$

181 The vector

182
$$\mathbf{y}_i = \left\{ \theta(t_i, \vec{x}_{\sigma_m(j)}) \right\}_{j=1}^m \quad (14)$$

183 collects the numerical observations of the water content. This vector is a nonlinear function \mathcal{H} of the state vector \mathbf{x}_i ,
184 described through the van Genuchten relation (10):

185
$$\mathbf{y}_i = \mathcal{H}(\mathbf{x}_i, \boldsymbol{\gamma}). \quad (15)$$

186 Equations (13) and (15) define the state-space model for the numerical simulations with CATHY.

187 **4. Estimation of saturated hydraulic conductivity**

188 The numerical simulation of LEO using CATHY with adequate initial and boundary conditions results in the
189 computation of θ on the entire domain Ω , and thus represents an alternative to the estimation of water content via
190 interpolation of the measurements. A key issue in the use of distributed models such as CATHY is the identification
191 of the model parameters that allow us to correctly describe the system and retrieve the measurements. In this context

192 the sensor network is fundamental for calibration of the numerical model and improving the forecast of the system
 193 state variables.

194 4.1. Data assimilation for inverse problems

195 The large number of sensors in LEO is useful in an inverse modeling framework, i.e., to infer the spatial distribu-
 196 tion of the soil hydraulic properties to be adopted in equations (9), (10), and (11). The inverse problem can be stated
 197 as follows: find the parameter vector, $\boldsymbol{\gamma}_p$, that minimizes the objective function given by the squared error between
 198 the observed and the simulated measurements:

$$199 \quad \boldsymbol{\gamma}^p = \arg \min_{\boldsymbol{\gamma} \in \mathcal{K}} \sum_{i=1}^{i_F} \|\mathbf{y}_i - \mathbf{y}_i^o\|^2 \quad (16)$$

200 where i_F is the total number of measurement times in the inversion experiment and \mathcal{K} is the search space where
 201 we look for the optimal solution. The numerical minimization of such an objective function is impractical for large
 202 state-space models and mainly depends on the dimension of the search space \mathcal{K} .

203 As an alternative, we consider sequential data assimilation methods (filtering) for parameter estimation [12, 30].
 204 Starting from a prior pdf of the parameters, $p_0(\boldsymbol{\gamma})$, the joint pdf of the state variables and the parameters is sequentially
 205 updated each time that an observation becomes available (assimilation time). Using the Bayes formula we obtain

$$206 \quad p^a(\mathbf{x}_i, \boldsymbol{\gamma}_i | \{\mathbf{y}_j\}_{j=1}^i) = C \mathcal{L}(\mathbf{y}_i | \mathbf{x}_i, \boldsymbol{\gamma}_i) p^f(\mathbf{x}_i, \boldsymbol{\gamma}_i | \{\mathbf{y}_j\}_{j=1}^{i-1}) \quad (17)$$

207 where p^a and p^f are the analysis and forecast pdfs, \mathcal{L} is the likelihood function, and C is a normalization constant. The
 208 forecast pdf, $p^f(\mathbf{x}_i, \boldsymbol{\gamma}_i | \{\mathbf{y}_j\}_{j=1}^{i-1})$, represents the evolution in time of the previous analysis, $p^a(\mathbf{x}_{i-1}, \boldsymbol{\gamma}_{i-1} | \{\mathbf{y}_j\}_{j=1}^{i-1})$, and
 209 is computed from the model equation (13) (see, e.g., [31]). The successive computation of the forecast and analysis
 210 pdfs represents the filtering problem. In the case of a linear state-space model with additive and Gaussian noise,
 211 the analysis and forecast pdfs are Gaussian and the Kalman filter [32] directly computes the expected values and the
 212 covariances of these distributions by minimising the variance of the analysis pdf.

213 For a nonlinear state-space model with update of both state vector and parameters, such as the one defined in (13)
 214 and (15), the filtering solution requires the use of random realizations to discretize the pdfs of interest and approximate
 215 their evolution in time.

216 4.2. Ensemble Kalman filter

217 In the EnKF method, the Kalman gain is still used to evaluate the analysis step and the relative expected values
 218 and covariance matrices. The optimality of the original Kalman filter (minimum variance) is lost and the Gaussian
 219 hypothesis describes the analysis pdf. The filter is initialized with an ensemble of N random samples of the prior
 220 distribution of the state, $\{\mathbf{x}_0^{a,j}\}_{j=1}^N \sim p(\mathbf{x}_0)$ and parameters $\{\boldsymbol{\gamma}_0^{a,j}\}_{j=1}^N \sim p(\boldsymbol{\gamma}_0)$. The empirical distribution of the numerical
 221 solutions $\{\mathbf{x}_i^{f,j}\}_{j=1}^N$,

$$222 \quad \mathbf{x}_i^{f,j} = \mathcal{F}(\mathbf{x}_{i-1}^{a,j}, q_{i-1}, \boldsymbol{\gamma}_{i-1}^{a,j}), \quad (18)$$

223 approximates the forecast step, while the parameters are assumed constant in the time interval $[t_{i-1}, t_i]$, i.e., $\boldsymbol{\gamma}_i^{f,j} = \boldsymbol{\gamma}_{i-1}^{a,j}$.
 224 The variable q_{i-1} represents the forcing term during the forecast step. In the analysis step of EnKF, both the state vector
 225 and the parameters are updated with the augmented state technique:

$$226 \quad \begin{pmatrix} \mathbf{x}_i^{a,j} \\ \boldsymbol{\gamma}_i^{a,j} \end{pmatrix} = \begin{pmatrix} \mathbf{x}_i^{f,j} \\ \boldsymbol{\gamma}_i^{f,j} \end{pmatrix} + \mathbf{K}_i^f (\mathbf{y}_i^{o,j} - \mathbf{y}_i^j) \quad (19)$$

227 where $\mathbf{y}_i^j = \mathcal{H}(\mathbf{x}_i^{f,j})$. The vector $\mathbf{y}_i^{o,j}$ represents random perturbations of the observed measurements \mathbf{y}_i^o according
 228 to (7). The use of the perturbed observations $\mathbf{y}_i^{o,j}$ theoretically guarantees that, in the Kalman filter hypothesis, the
 229 correct variance of the updated variables will be computed (see, e.g., [33]). \mathbf{K}_i^f is an ensemble approximation of the
 230 Kalman filter taking into account the cross correlations between the augmented state and the observations (for more
 231 details see, e.g., [27] and [29]). The empirical distribution of the parameters obtained at the last assimilation time,
 232 $\{\boldsymbol{\gamma}_{i_f}^{a,j}\}$, represents the posterior distribution of the parameters, which depends on the number of active sensors m and
 233 their location in the domain Ω .

234 The inverse methodology presented in this section employs the measurements at the active sensors to estimate
 235 the posterior distribution of the hydraulic conductivity field at LEO. Decreasing the number of active sensors might
 236 result in high uncertainty associated with the parameters or in unrealistic parameter estimates. In a synthetic scenario
 237 we can directly compare the true and the posterior distributions of the conductivity field. Since this approach is not
 238 possible in real applications, to understand if the number of active sensors is sufficient to obtain an accurate posterior
 239 distribution of the parameters, we re-run the simulation (without assimilation) employing the posterior distribution of
 240 the parameters. If the numerical outputs are inside the confidence interval of the measurements, then we can conclude

241 that the estimated parameters are admissible. In this case we can say that the number of active sensors is sufficient for
 242 the calibration of the model and for the observability of the system.

243 5. Numerical experiments

244 For the estimation of water content and the retrieval of the saturated conductivity we consider only synthetic
 245 measurements generated with the numerical model CATHY. This allows evaluation of the accuracy and robustness of
 246 the proposed identification approaches. We describe below the setup of the numerical experiment used to obtain the
 247 measurement set.

248 The LEO hillslope domain Ω is discretized with $22 \times 60 = 1320$ square cells of dimension 0.5×0.5 m at the surface
 249 and 20 vertical layers of thickness of 0.05 m. The obtained parallelepipeds are further subdivided into six tetrahedra
 250 for the FEM solution of Eq. (9) with piecewise linear basis functions. The numerical simulations reproduce the first
 251 LEO experiment, conducted on February 18, 2013. A rainfall rate of 12 mm/h is imposed for a duration of 22 h
 252 followed by an evaporative forcing of 1 mm/d until the end of the experiment. Gevaert et al. [34] presents a detailed
 253 description of the experimental results. This experiment is a suitable platform for testing the reliability of the sensor
 254 network since the hillslope undergoes a broad range of dynamics and system states from fully unsaturated conditions
 255 to significant surface and subsurface outflows. Niu et al. [10] were able to reproduce the overland flow, seepage face
 256 flow, and total water storage responses with the CATHY model and a heterogeneous configuration of soil parameters.

257 *Generation of the synthetic truth.* We assume that the true saturated hydraulic conductivity field K_S is a realization
 258 of a stationary three-dimensional random field with a lognormal distribution,

$$259 \quad Y = \text{Log}(K_S), \quad Y \sim N(\mu_Y, C_Y), \quad (20)$$

260 and exponential covariance function, C_Y , with vertical anisotropy:

$$261 \quad C_Y(l_x, l_y, l_z) = \sigma_Y \exp \left(- \sqrt{ \left(\frac{l_x}{\lambda_x} \right)^2 + \left(\frac{l_y}{\lambda_y} \right)^2 + \left(\frac{l_z}{\lambda_z} \right)^2 } \right) \quad (21)$$

262 where l_x , l_y , and l_z are the lag distances and λ_x , λ_y , and λ_z are the integral scales in direction x , y , and z , respectively.
 263 μ_Y and σ_Y are the expected value and variance of Y . The other parameters describing the soil hydraulic properties

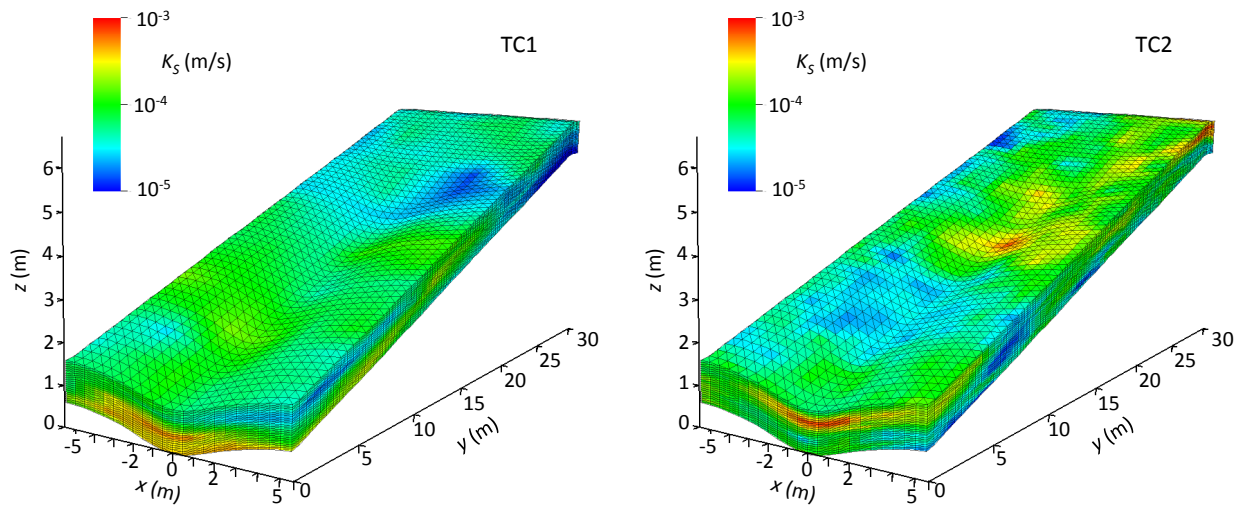


Figure 3: Spatial distribution of the hydraulic conductivity K_S in the domain Ω for the true runs of TC1 and TC2.

264 are considered homogeneous, and Table 1 summarizes their values. The parameters μ_Y and σ_Y are calculated using
 265 an expected value of K_S equal to 1.0×10^{-4} m/s, in agreement with the calibration proposed in [10], and a coefficient
 266 of variation of K_S equal to 100%, corresponding to variations of K_S of two orders of magnitude. The resulting
 267 parameters of the Gaussian field Y are $\mu_Y = -9.56 \log(\text{m/s})$ and $\sigma_Y = 0.83 \log(\text{m/s})$. To assess the sensitivity of the
 268 results with respect to variations of the Y integral scale, which is unknown for the real LEO soil, we consider two
 269 scenarios: in test case 1 (TC1) we set $\lambda_x = \lambda_y = 8$ m and $\lambda_z = 0.5$ m whereas in test case 2 (TC2) the integral scales
 270 are halved, $\lambda_x = \lambda_y = 4$ m and $\lambda_z = 0.25$ m. A 3-dimensional adaptation of the random generator of stochastic fields
 271 HYDRO_GEN [35, 36] is employed to sample the realizations of the log-conductivity Y .

272 Figure 3 shows the true hydraulic conductivity fields that characterize test cases TC1 and TC2. The hydrological
 273 response associated with these two configurations of hydraulic conductivity is represented by the black straight lines
 274 shown in Figure 4 denoted as “true” state. Starting from unsaturated initial conditions ($V_{\Omega}(t_0) \approx 35 \text{ m}^3$), the rain water
 275 infiltrates in the soil and accumulates at the base of Ω , increasing the water storage. At around $t = 10$ h the formation
 276 of base flow generates an outgoing flux at the seepage face. Overland flow appears at the outlet after 20 h. From a
 277 qualitative point of view, these results are in accordance with the observations and simulations presented in [10]. The
 278 intensity and timing of the seepage face and overland flows are different in TC1 and TC2, due to the different soil
 279 hydraulic properties.

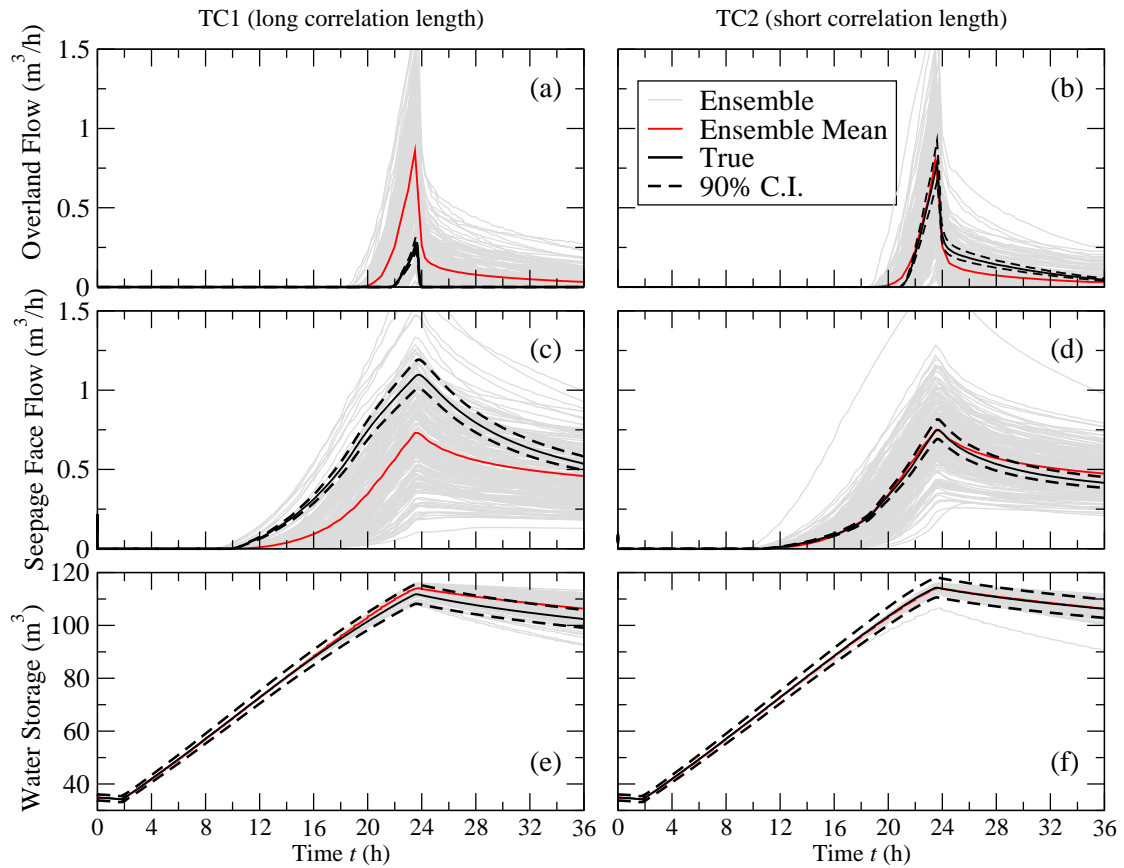


Figure 4: Hydrological response of the CATHY model for test cases TC1 and TC2. The grey lines represent the outputs obtained with 200 random realizations of the prior pdf of the saturated hydraulic conductivity and the red line is their mean. The solid black line is the response associated with the true run, and the dashed lines delimit the 90% confidence interval of the true measurements.

Grid information	
DEM cell dimensions	0.5×0.5 m
Number of cells in the surface grid	$22 \times 60 = 1320$
Soil depth (uniform)	1 m
Vertical discretization (# of layers)	20
Soil layer thickness (uniform)	0.05 m
# of nodes in the 3D grid	$23 \times 61 \times 21 = 29463$
# of tetrahedral elements in the 3D grid	158400
Parameters	
Saturated hydraulic conductivity	Heterogeneous lognormally distributed $E[K_S] = 1.0 \times 10^{-4}$ m/s, $CV_{K_S} = 100\%$ $Y = \log(K_S)$ has exponential covariance function TC1: $\lambda_x = \lambda_y = 8$ m; $\lambda_z = 0.50$ m TC2: $\lambda_x = \lambda_y = 4$ m; $\lambda_z = 0.25$ m
Parameters	
Aquifer specific storage	Homogeneous $S_s = 5 \times 10^{-4}$ m ⁻¹
Porosity	$\theta_s = 0.37$
Van Genuchten curve fitting parameters	$\alpha = 0.6$ m ⁻¹ , $n = 2.26$
Residual moisture content	$\theta_r = 0.002$
Simulation period	
129600 s (36 h)	
Initial conditions	
Linear interpolation of LEO water content sensors at 8:00 a.m. 02/18/2013	
Atmospheric forcing	
Spatially distributed rainfall (12 mm h ⁻¹) and evapotranspiration (0.04 mm h ⁻¹)	
Measurements	
Measures of volumetric water content θ	normally distributed with $\sigma_\theta \approx 0.012$
Measures of storage V_Ω	lognormally distributed with $cv_{V_\Omega} = 2\%$
Measures of seepage face volumes V_T	lognormally distributed with $cv_{V_T} = 2\%$
Ensemble size N	200

Table 1: Domain discretization and parameter values for the LEO sensor failure experiment. cv indicates the coefficient of variation.

280 *Generation of the open loop.* The open loop (OL) represents the sensitivity of the model response to variations of the
281 parameters. We evaluate OL from a set of model solutions constructed from 200 independent Y realizations sampled
282 from the prior pdf described above. The OL results are shown in Figure 4. Note that the response of the true realization
283 may differ drastically from the response of the OL ensemble mean (red line), as occurring in TC1. In the TC2 case
284 the opposite behaviour is observed, but this is just a random result. In both scenarios, the spread of the OL response
285 reveals a large uncertainty in the model forecast, suggesting that the ensemble mean is not a statistically accurate
286 estimator of the true response.

287 *Generation of synthetic observations.* The water content measurements employed in the interpolation and inverse
288 problems are selected from the results of the true runs in TC1 and TC2. To simulate the measurement error, the values
289 of water content are perturbed according to equation (7).

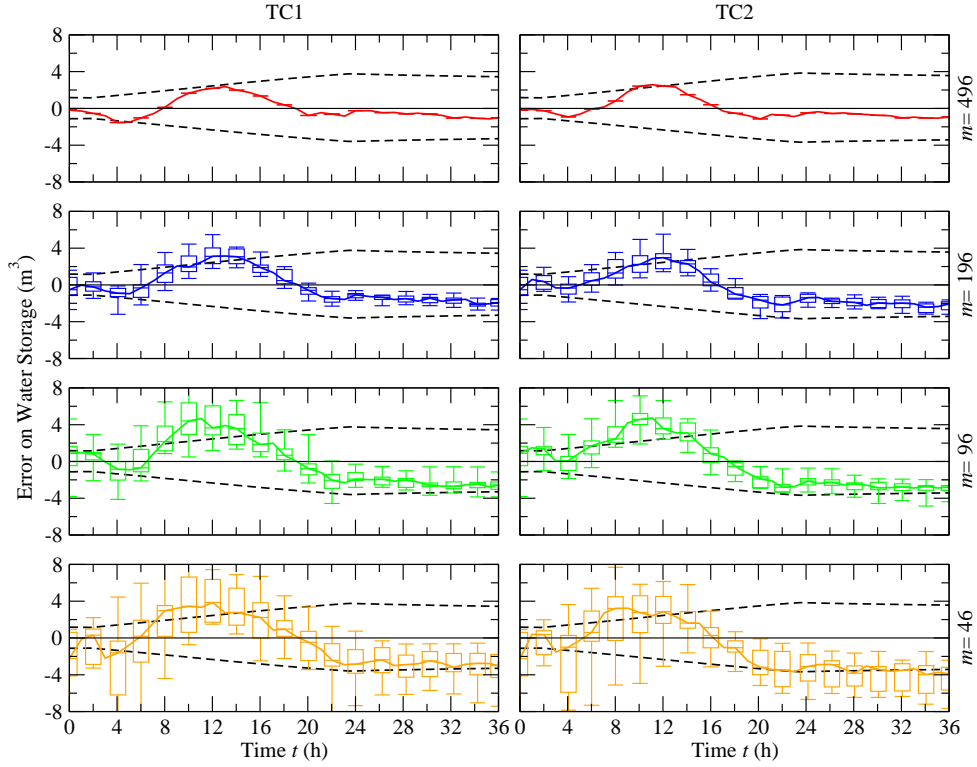


Figure 5: Error between the measured water storage in Ω and the one estimated with $\hat{\theta}$ (equation (8)). The dashed lines represent the 90% confidence intervals, while the box plots indicate the variability of the results with respect to the spatial distribution of the sensors.

290 6. Results and discussion

291 6.1. Reconstruction of volumetric water content

292 In this section the perturbed measurements of volumetric water content are employed to estimate the total water
 293 volume in the entire domain Ω , as described in equation (8). The sensor failure analysis is performed considering
 294 a decreasing number of active sensors, $m=[496, 446, 396, 346, 296, 246, 196, 146, 96, 46, 21]$, and $r=10$ random
 295 distributions of the active sensors in space.

296 Figure 5 presents the time behaviour of the error between the measured volumes (shown as the solid black line in
 297 Figure 4, panels (e) and (f)) and the estimated volumes with m sensors, $m=496, 196, 96$, and 46 . The box plots for the
 298 $m=496, 196, 96$, and 46 represent the 2.5, 25, 50, 75, and 97.5 percentiles of the errors obtained with $r=10$ different
 299 spatial configurations of the active sensors. The estimation of water storage using the entire sensor network ($m=496$)
 300 produces errors bounded by the 90% confidence interval (dashed lines), showing the accuracy of this estimate. The

301 largest errors are recorded during the infiltration period, $2 \text{ h} < t < 12 \text{ h}$, i.e., when the soil is still mostly unsaturated
 302 and the linear interpolation is not optimal in describing the nonlinear distribution of the infiltration front. A better
 303 estimation of V_Ω is obtained for $t > 20 \text{ h}$, i.e., when the domain is almost completely saturated. In fact, in this scenario
 304 the distribution of water content is more homogeneous, especially at the base of LEO (where sensors are sparser)
 305 resulting in more accuracy of the linear interpolation/extrapolation. Similar water volumes are estimated interpolating
 306 the measurements from less than the half of the sensors, $m=196$. The location of the active sensors slightly influences
 307 the estimation of water storage, especially during the infiltration period, i.e., when a large number of measurements is
 308 required to capture the water distribution in Ω . In such a period, some spatial configurations of the sensors generate
 309 an error that exceeds the limits of the 90% confidence interval, although all the placements are able to reproduce the
 310 correct volumes with sufficient accuracy during the rest of the simulation. When the number of sensors drops below
 311 100 ($m=96$, $m=46$), the estimation of water storage becomes less reliable and highly dependent on the distribution
 312 of the active sensors. The results of the smaller sensor sets show unreliable estimates during most of the simulation.
 313 These results are not sensitive to the integral scale of the hydraulic conductivity.

314 Since the measurements are generated from a numerical simulation, we can compute the spatial behaviour of the
 315 error between the estimated and the true distribution of volumetric water content in Ω . Panel (a) in Figure 6 shows the
 316 spatial distribution of volumetric water content at time $t=12 \text{ h}$ for TC2. As expected, at this time the uppermost layers
 317 of the landscape are partially saturated, while base flow starts forming at the bottom of the domain. Panels (b) and (c)
 318 show the errors for the $m=496$ and $m=46$ cases of test TC2, respectively. It is evident that the error is smaller where
 319 the density of observations is higher.

320 To summarize the results for all the numerical simulations, we compute the time average of the L_2 -norm of the
 321 error between the estimated and the true water content,

$$322 \quad e_\theta = \frac{1}{i_F} \sum_{i=1}^{i_F} \left[\int_{\Omega} (\tilde{\theta}(t_i) - \theta(t_i))^2 d\Omega \right]^{\frac{1}{2}} \quad (22)$$

323 Figure 7 shows the values of e_θ for all the combinations and number of active sensors considered. As expected, the
 324 water content estimation with $m=496$ sensors has the minimum error. The error increases and is more sensitive to the
 325 location of the sensors as the number of active sensors decreases. In particular the interpolation errors obtained with
 326 $m < 100$ active sensors are consistently higher than the errors obtained with $m > 300$, meaning that there is a strong

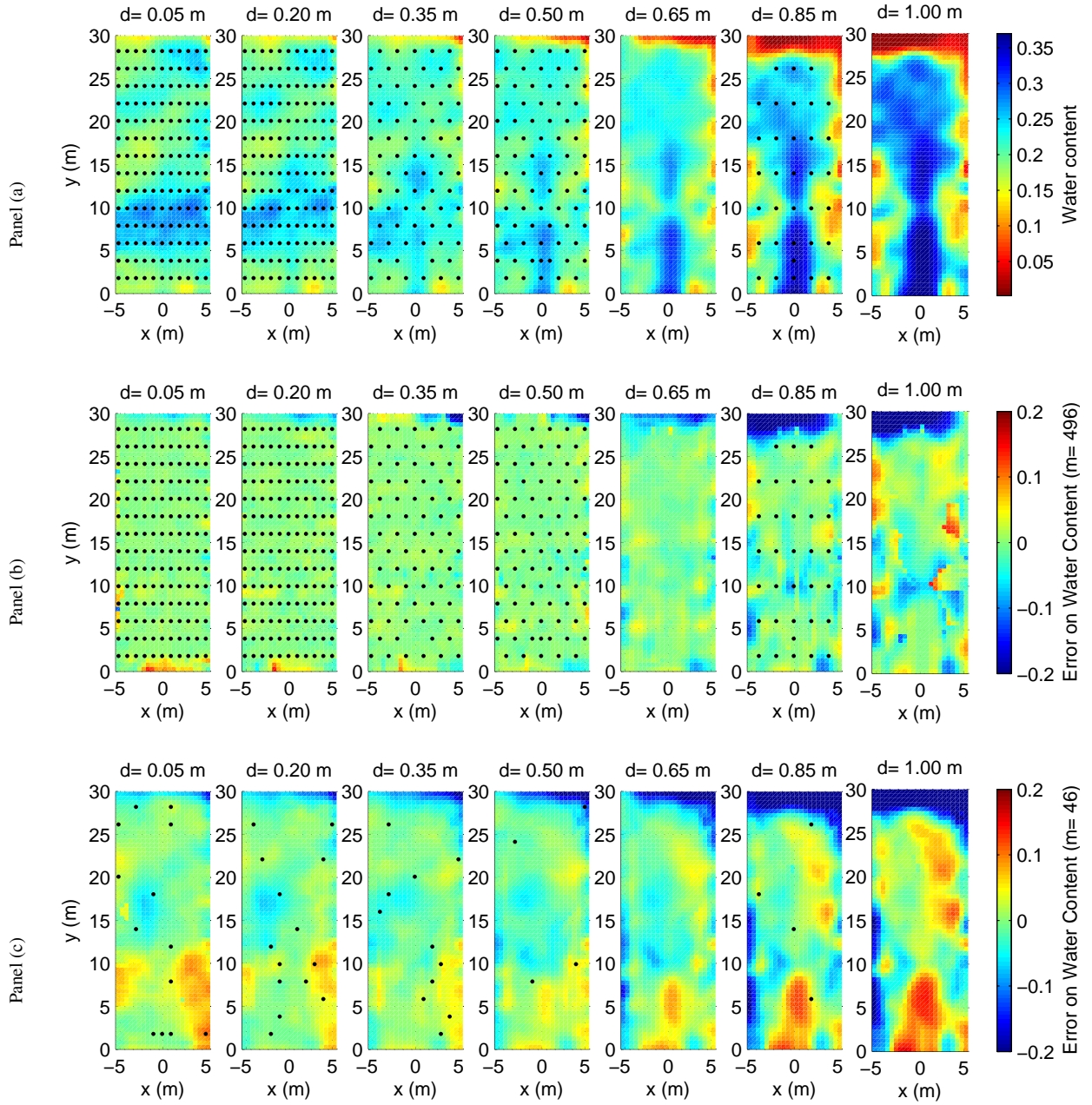


Figure 6: True volumetric water content for test TC2 after 12 h at seven depths (a) and the errors associated with the estimated volumetric water content interpolating the measurements from $m=496$ (b) and $m=46$ (c) active sensors. The black dots show the locations of the active sensors.

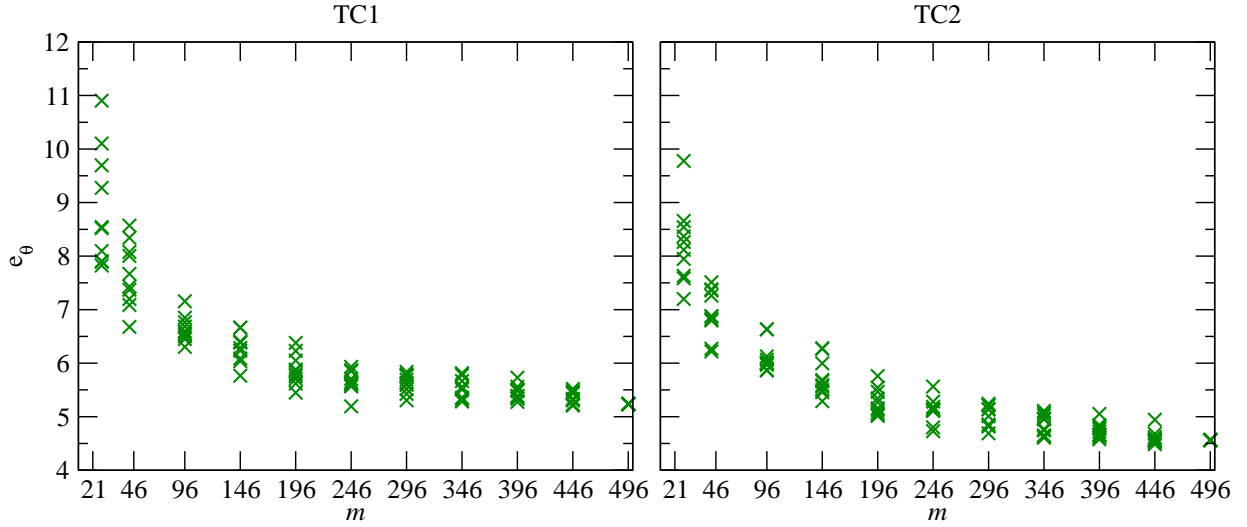


Figure 7: Time-averaged L^2 -norm of the error between the true and estimated water contents (equation (22)) for the different number and configurations of active sensors.

327 deterioration in the observability of the system.

328 6.2. Reconstruction of hydraulic conductivity

329 In Figure 4 we show that the response of the hydrological model CATHY is highly sensitive to the different
330 samples of the hydraulic conductivity field. It is evident that the prior distribution of K_S entails high uncertainty in
331 the model results, implying that the model, with an incorrect parameterization, is not suitable for the prediction of the
332 water dynamics in the landscape. Here we assimilate the measurements of volumetric water content using EnKF to
333 obtain a posterior distribution of the hydraulic conductivity field that is closer to the realization adopted in the true
334 runs. The sensor failure analysis is performed considering a decreasing number of active sensors, $m=[496, 196, 146,$
335 $96, 46, 21]$, and one random distribution of the active sensors in space ($r=1$).

336 At every assimilation time EnKF updates the model state variables (i.e., pressure head and, if present, surface dis-
337 charge) and the hydraulic conductivity field associated with each MC realization. Thus, we expect that the discrepancy
338 between the estimated log-conductivity field, \tilde{Y} , and the true field, Y , decreases during the filtering process. We con-
339 sider the root mean squared error between the ensemble of log-conductivity realizations and the true log-conductivity,

$$340 \quad e_Y(t) = E \left[\frac{1}{\Omega} \int_{\Omega} (\tilde{Y}(t) - Y) d\Omega \right] \quad (23)$$

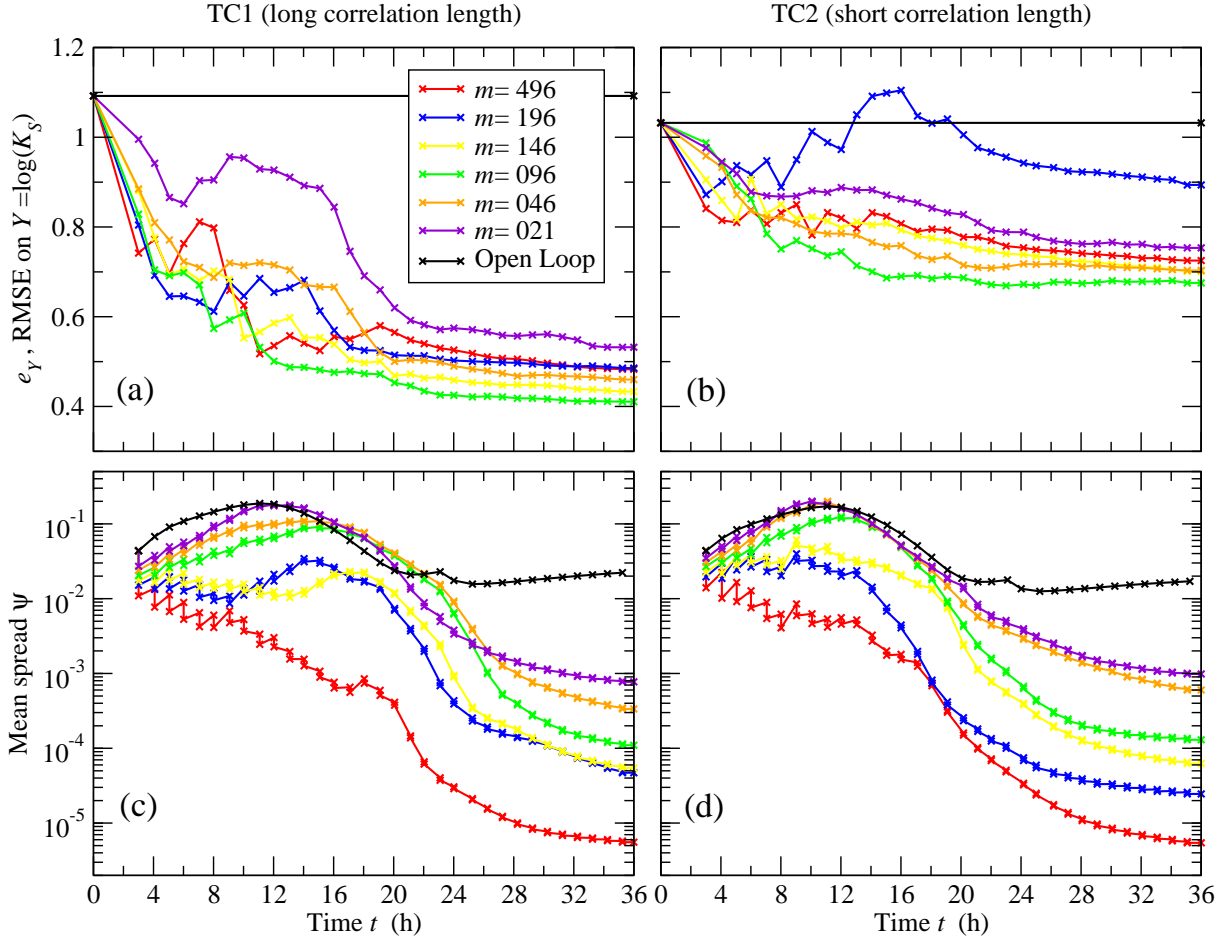


Figure 8: Root mean squared error (equation (23)) between the true and estimated distributions of K_S (panels (a) and (b)) and mean ensemble spread (equation (24)) of the pressure heads (panels (c) and (d)) at each assimilation of EnKF for different numbers of active sensors. The first update is at $t=3$ h.

341 where $\tilde{Y}(t)$ is the log-conductivity random field described by the MC realizations at time t .

342 The values of $e_Y(t)$ at the assimilation times are shown in panels (a) and (b) of Figure 8 for TC1 and TC2, re-
 343 spectively. We can see that even for a low number of active sensors the assimilation process computes a posterior
 344 distribution of \tilde{Y} that has a lower error than the prior. The first assimilation step produces the most significant cor-
 345 rection on the ensemble of \tilde{Y} and, at this assimilation time, there is a clear correspondence between the simulations
 346 with the smallest error e_Y and the simulations with the largest number of active sensors m . However, the errors do not
 347 show a monotone decrease in time. This undesired result, that would be expected when using a low number of sensors

348 (e.g., $m=21$), is particularly affecting the simulations with a large number of measurements (e.g., $m=196$, $m=496$),
 349 resulting in a final distribution of the errors that is not consistent with the number of active sensors. We associate
 350 this behaviour to two main drawbacks of EnKF. On the one hand, EnKF performs a global update of the parameters
 351 at each assimilation time. The ensemble approximation of the cross covariances between the state variables and the
 352 parameters may result in spurious correlations and a (wrong) update of parameters that are not directly involved in the
 353 measured process. On the other hand, EnKF tends to underestimate the variance of the updated state variables, thus
 354 reducing the efficiency of the Kalman gain in correcting the erroneous updates. This results in the reduction of the
 355 ensemble spread, especially for the scenarios with a large number of sensors. To show this phenomenon, we evaluate
 356 the mean variance associated with the nodal pressure head, which is the main state variable of the CATHY model and
 357 is directly related to the volumetric water content, as a measure of the ensemble spread:

$$358 \quad \bar{\sigma}_{\psi}^2(t) = \frac{1}{\Omega} \int_{\Omega} E [(\psi(t) - E[\psi(t)])^2] d\Omega \quad (24)$$

359 where E indicates the mean over the ensemble. Panels (c) and (d) of Figure 8 show the temporal value of $\bar{\sigma}_{\psi}^2(t)$ for
 360 TC1 and TC2, respectively. We can see that, in all the scenarios explored, the variance decreases with respect to the
 361 OL run, indicative of a lower uncertainty on the model response. It is evident that the simulations with a high number
 362 of sensors have a faster decrease in the ensemble spread.

363 In our experiment, the infiltration process at the first assimilation time ($t = 3$ h) involves only the superficial
 364 layers of the domain. Thus the first assimilation steps are able to correct the Y field at the surface but have less
 365 accuracy at the bottom layers of the domain. As a consequence, the following update steps are necessary to improve
 366 the parameter estimation for these layers. Common techniques adopted to improve the EnKF update are covariance
 367 localization [21], to remove spurious correlations, and covariance inflation [19], to increase some components of
 368 the covariance and improve the ensemble spread. Here we adopt a different approach. The idea is to increase the
 369 forecast time between assimilations, in such a way as to have a better correlation between the state realizations and
 370 the associated parameters. Firstly, we focus on the time of the first assimilation, which entails the maximum decrease
 371 in the error and in the ensemble spread. It is important to perform the first update when the infiltrating water has also
 372 reached the sensors at the deepest layer, so that the process dynamics influences the response of the sensors throughout
 373 the entire domain. In our simulations, this occurs at $t=12$ h. As shown in panels (c) and (d) of Figure 8, this time

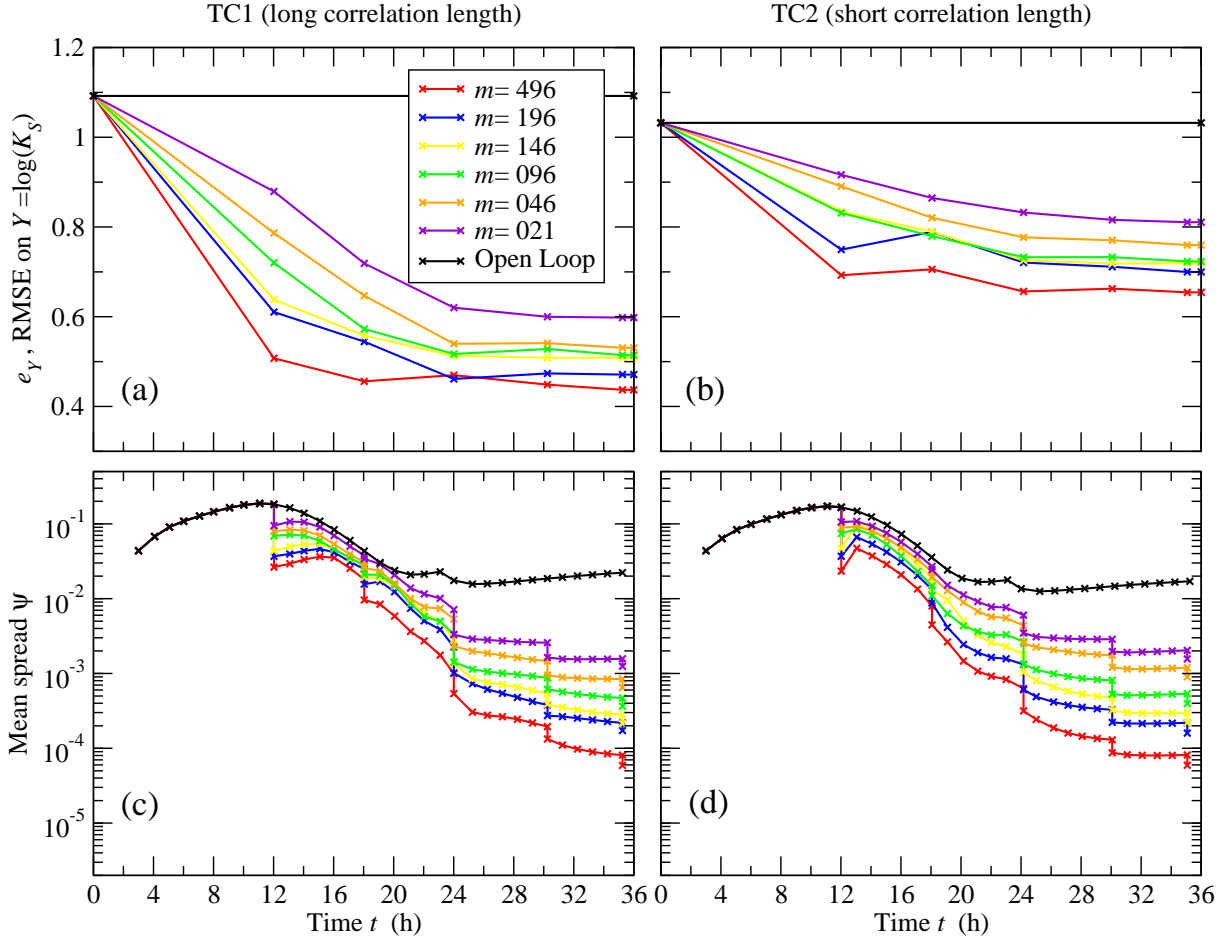


Figure 9: Root mean squared error (equation (23)) between the true and estimated distributions of K_S (panels (a) and (b)) and mean ensemble spread (equation (24)) of the pressure heads (panels (c) and (d)) at each assimilation of EnKF for different numbers of active sensors. The first update is at $t=12$ h and the assimilation frequency is six hours.

374 corresponds to the maximum variance on the OL pressure heads, and thus to the maximum differentiation between
 375 the ensemble realizations obtained with the prior parameter distribution. Secondly, as Shi et al. [37] have pointed out,
 376 the accuracy of an EnKF analysis may degrade when using short assimilation intervals, thus we increased the update
 377 time from hourly to every 6 h.

378 Figure 9 shows the convergence profiles of the errors e_Y obtained with the first assimilation timed at $t=12$ h and
 379 an assimilation frequency of six hours. It can be seen that the errors at the first update are lower than in the previous
 380 scenario (Figure 8) for all sensor failure configurations explored. Moreover, the error now increases only at a small

381 number of assimilation times, very slightly, and without compromising the final correlation between the errors and
382 the number of active sensors. The greatly reduced number of updates owing to the postponed first update and larger
383 assimilation interval may explain the higher final errors obtained with fewer active sensors. However, the higher values
384 of the ensemble spread should allow improved performance for longer simulations, thus restoring (or improving) the
385 previously computed errors on the parameters.

386 Figure 9 shows that, for both TC1 and TC2, the error at the end of the simulation increases consistently as the
387 number of active sensors drops. This suggests that the hydrological dynamics of LEO is well captured by the sensor
388 network even when a large fraction of sensors has failed.

389 Figures 10 and 11 compare the true conductivity field and the a posteriori ensemble mean of the conductivity field
390 estimated with $m=496$ and $m=21$ active sensors at six different depths d for TC1 and TC2, respectively. For both test
391 cases, the estimated conductivity with $m=496$ captures most of the features of the true field, with small differences at
392 the bottom layer where there are no sensors. The final K_S estimated with $m=21$, although having larger errors (see
393 Figure 9), manages to capture several zones of high and low conductivity of the true field, especially for the scenario
394 with long correlation length (TC1). This means that, also with only 21 active sensors, EnKF is able to improve the
395 distribution representing the true soil hydraulic conductivity with respect to the a priori pdf.

396 This is even more evident in Figure 12, where the final ensemble means of the conductivities are employed in a
397 forward run of CATHY to compare the hydrological response of the system for the different distributions of K_S . The
398 true K_S and the K_S estimated with $m=496$, 46, and 21 active sensors produce very similar integrated responses of the
399 system in terms of overland flow (panels (a) and (b)), seepage face flow (panels (c) and (d)), and total water storage
400 (panels (e) and (f)) for both TC1 and TC2. The differences in the soil conductivity fields have a noticeable impact only
401 when we examine the water content measured by the sensor network. Panels (g) and (h) report the root mean squared
402 error on the measured volumetric water content, and we see that, for both TC1 and TC2, the results with $m=496$ active
403 sensors are more accurate than those using $m=21$ or $m=46$ sensors.

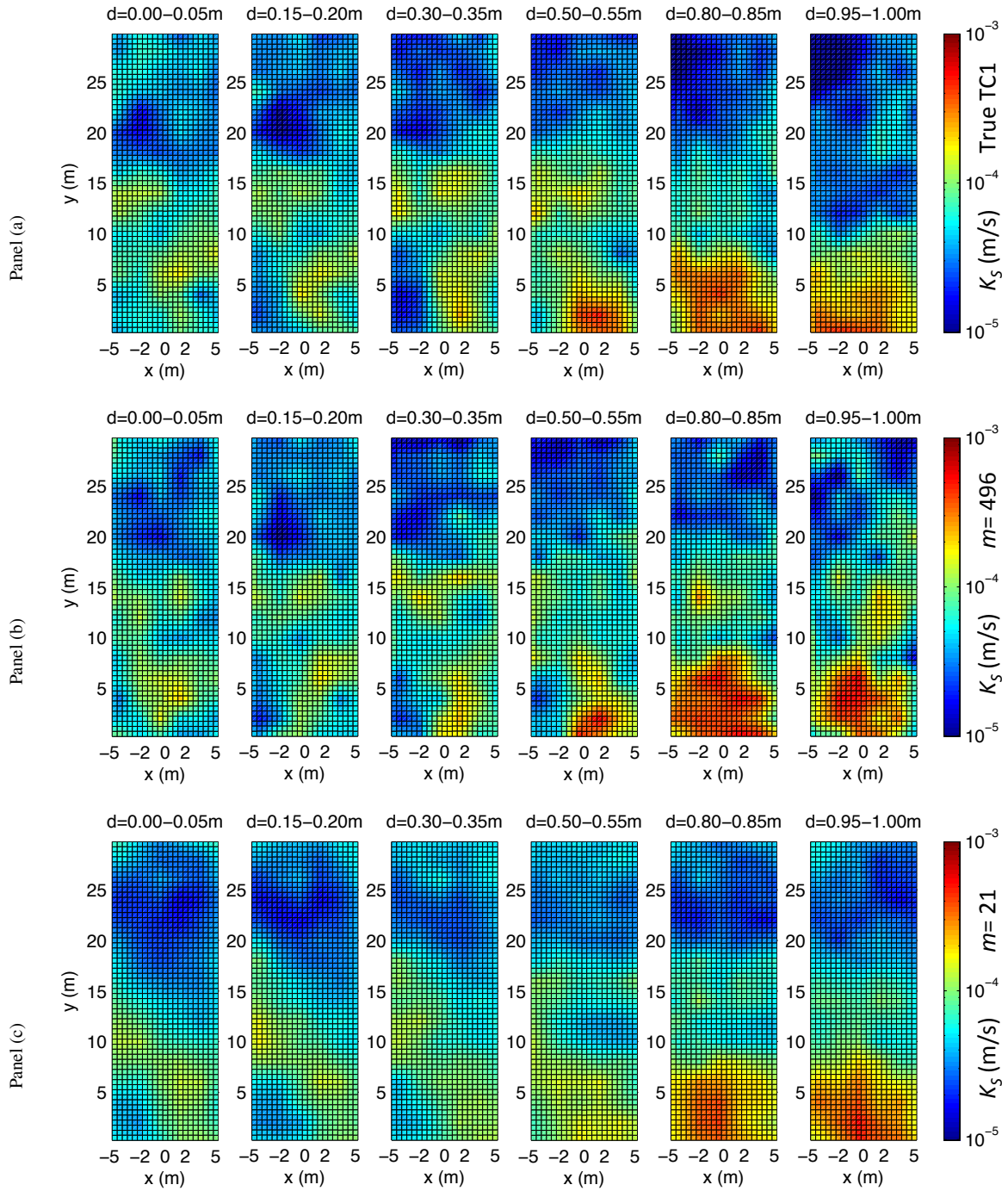


Figure 10: Test case TC1. Comparison between the true distribution of K_S (a) and the ensemble mean of K_S estimated with the EnKF procedure using $m=496$ (b) and $m=21$ (c) active sensors.

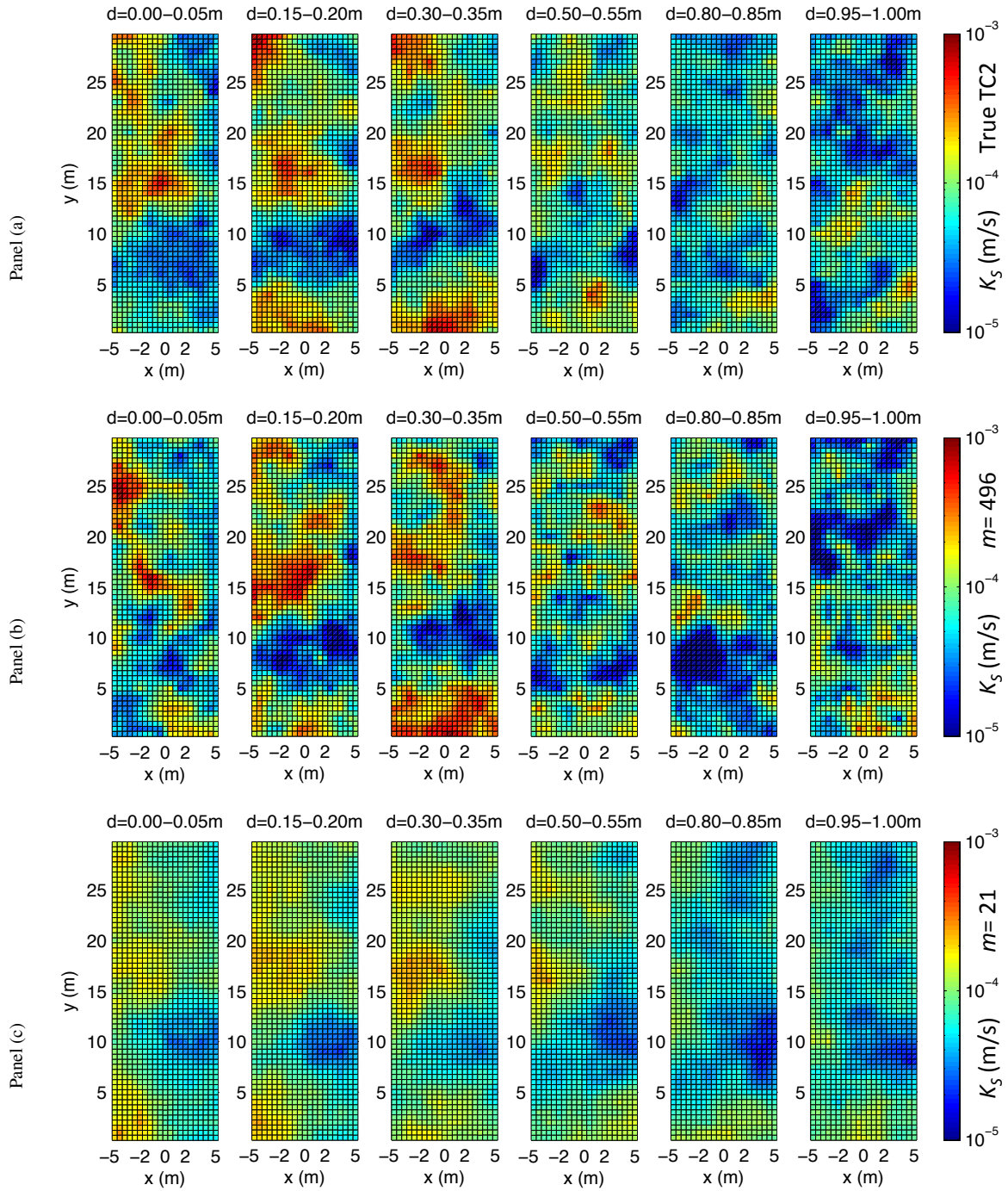


Figure 11: Test case TC2. Comparison between the true distribution of K_S (a) and the ensemble mean of K_S estimated with the EnKF procedure using $m=496$ (b) and $m=21$ (c) active sensors.

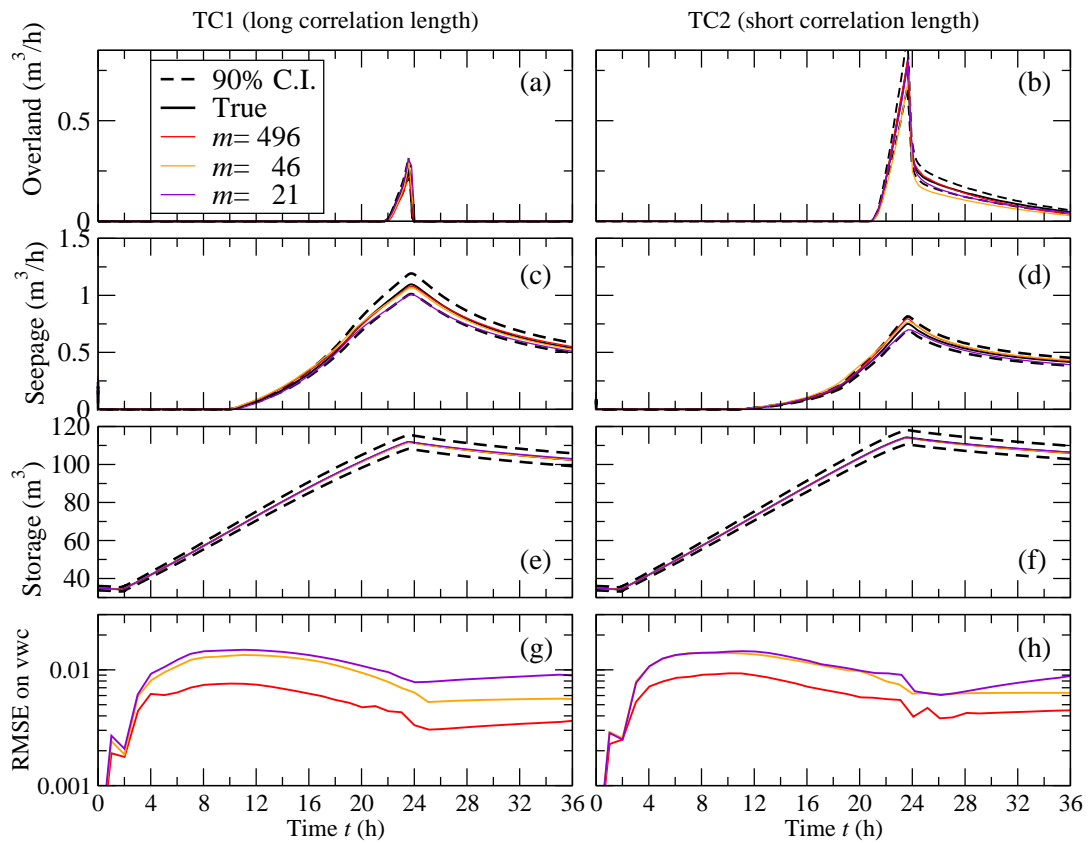


Figure 12: Comparison between the hydrological responses of CATHY using the true K_S field and the estimated distribution of K_S for $m=496$, 46, and 21 active sensors, for both the TC1 (left graphs) and TC2 (right graphs) test cases. The hydrological responses shown are, from top graphs to bottom, overland flow, seepage face flow, total water storage, and root mean squared error on the volumetric water content. C.I. is confidence interval.

404 7. Conclusions

405 We have presented a failure analysis of the sensor network deployed in the experimental hillslopes of the Land-
406 scape Evolution Observatory at Biosphere 2. The main objective of this study was the determination of the minimum
407 number of active sensors that are necessary for a reliable observation of the water dynamics in the system. To reach
408 this goal we elected to work with two general quantities of interest in hydrological modeling. The first quantity is
409 the total water volume inside the hillslope at a given time. The second quantity is the total discharge at the outlet, as
410 formed by the sum of the seepage and overland fluxes. Synthetic test cases were developed for 11 different sensors
411 sets having numbers of active sensors varying between 496 and 21. The soil was modeled as a three-dimensional
412 second order stationary random field with exponential covariance function. A simulation with a given K_s distribution
413 was used to obtain a “true” hillslope behavior, as simulated by the CATHY model, from which appropriate synthetic
414 measurements were taken. For the first quantity, the assessment of the sensor failure was carried out by comparing
415 the temporal variation of the total water storage as measured by virtual load cells and the total water volume obtained
416 by linear interpolation of the observed values at the spatially distributed water content sensors. In the case of outlet
417 discharge, the sensor network reliability was tested by using the measured water content inside the domain to identify
418 via EnKF the spatially heterogeneous saturated hydraulic conductivity. The accuracy of the identification was verified
419 by simulating the event with the hydrological model CATHY and comparing simulated and observed seepage and
420 overland flow.

421 The results showed that the reconstruction of volumetric water content via linear interpolation of the sensor net-
422 work values is an accurate procedure when the level of failure is low. When only few sensors are active (i.e., $m < 100$)
423 this methodology may introduce sizeable errors in the estimation of total water storage, especially during water in-
424 filtration periods. This corresponds to the fact that when the sensor network is only partially active there are large
425 portions of the hillslope where information is lost and interpolation identifies incorrect water content values, especially
426 near the soil surface.

427 Retrieval of the saturated hydraulic conductivity distribution, on the other hand, seems to be more robust. In fact,
428 calibration of CATHY via EnKF is able to improve the estimation of the true conductivity fields yielding accurate
429 predictions of the global hydrological response of the LEO hillslope also at the lowest tested numbers of active

430 sensors ($m=46$ and $m=21$). This result is achieved thanks to the role played by the hydrological model in correlating
431 in space and time the water content measurements with the hillslope response, and by the assumption of no model
432 bias that is inherent in our synthetic experiment. To thoroughly verify this result, we conducted two test cases where
433 the “true” saturated conductivity distribution was characterized by different correlation lengths. For both test cases we
434 were able to assert a high reliability of the sensor network. The CATHY simulations were able to accurately simulate
435 the behavior of the system also for the lowest number of active sensors. We remark that this was true for the overall
436 global response of the system, i.e., by looking only at total seepage and overland flow at the outlet. The recovery of the
437 distribution of the internal states was much more inaccurate especially when few sensors are active. In this case, the
438 simulations suffer from the problem that the EnKF identification of K_s yields a field that is statistically equivalent (up
439 to small errors) to the true realization. Hence, local details of the system dynamics, in terms of spatially distributed
440 water content, are lost, leading to inaccurate simulated water content values at the sensor positions, with errors that
441 decrease drastically when the entire sensor network is considered active. The outlet discharge, on the other hand, does
442 not suffer from this problem, a sign that the statistical properties (mean and variance) of the K_s field were correctly
443 identified.

444 The results from the two test cases showed that the identification process is more accurate for the larger correlation
445 length, corresponding intuitively to a lower number of parameters to be identified. Moreover, to improve the parameter
446 identification and limit the effects of erratic covariance estimations in the Kalman gain in areas of the domain where the
447 measurements contribute negligible information, we showed that decreasing the assimilation frequency and delaying
448 the first update helps incorporate responses to the infiltration signal from all the sensors. Consequently the sample
449 spread increases, yielding reduced inconsistencies between filter behavior and number of active sensors. In this case,
450 however, achieving small estimation errors requires larger simulation times, corresponding to a larger number of
451 update steps. The test case results also showed differences in reconstruction according to the degree of sensor failure.
452 Nonetheless we are able to reproduce the integrated hydrological response of LEO also with the lowest numbers of
453 active sensors.

454 Acknowledgments

455 The Biosphere 2 facility and the capital required to conceive and construct LEO were provided through a charitable
456 donation from the Phileology Foundation, and its founder, Mr. Edward Bass. We gratefully acknowledge this
457 charitable donation.

458 References

- 459 [1] N-S. Hsu and W. W-G. Yeh. Optimum experimental design for parameter identification in groundwater hydrology. *Water Resour. Res.*, 25:
460 1025–1040, 1989. doi: 10.1029/WR025i005p01025.
- 461 [2] T. T. Ushijima and W. W-G. Yeh. Experimental design for estimating unknown groundwater pumping using genetic algorithm and reduced
462 order model. *Water Resour. Res.*, 49(10):6688–6699, 2013. doi: 10.1002/wrcr.20513.
- 463 [3] L. Hopp, C. Harman, S. L. E. Desilets, C. B. Graham, J. J. McDonnell, and P. A. Troch. Hillslope hydrology under glass: confronting
464 fundamental questions of soil-water-biota coevolution at Biosphere 2. *Hydrol. Earth Syst. Sci.*, 13:2105–2118, 2009. doi: 10.5194/hess-13-
465 2105-2009.
- 466 [4] K. Dontsova, C. I. Steefel, S. Desilets, A. Thompson, and J. Chorover. Solid phase evolution in the Biosphere 2 hillslope experiment as
467 predicted by modeling of hydrologic and geochemical fluxes. *Hydrol. Earth Syst. Sci.*, 13:2273–2286, 2009. doi: 10.5194/hess-13-2273-
468 2009.
- 469 [5] V. Y. Ivanov, S. Fatichi, G. D. Jenerette, J. F. Espeleta, P. A. Troch, and T. E. Huxman. Hysteresis of soil moisture spatial heterogeneity and
470 the “homogenizing” effect of vegetation. *Water Resour. Res.*, 46:W09521, 2010. doi: 10.1029/2009WR008611.
- 471 [6] M. Camporese, C. Paniconi, M. Putti, and S. Orlandini. Surface-subsurface flow modeling with path-based runoff routing, bound-
472 ary condition-based coupling, and assimilation of multisource observation data. *Water Resour. Res.*, 46:W02512, 2010. doi:
473 10.1029/2008WR007536.
- 474 [7] C. Paniconi and M. Putti. A comparison of Picard and Newton iteration in the numerical-solution of multidimensional variably saturated flow
475 problems. *Water Resour. Res.*, 30(12):3357–3374, 1994. doi: 10.1029/94WR02046.
- 476 [8] S. Orlandini and R. Rosso. Diffusion wave modeling of distributed catchment dynamics. *J. Hydraul. Eng.-ASCE*, 1(3):103–113, 1996. doi:
477 10.1061/(ASCE)1084-0699(1996)1:3(103).
- 478 [9] S. Orlandini, G. Moretti, M. Franchini, B. Aldighieri, and B. Testa. Path-based methods for the determination of nondispersive drainage
479 directions in grid-based digital elevation models. *Water Resour. Res.*, 39(6):1144, 2003. doi: 10.1029/2002WR001639.
- 480 [10] G.-Y. Niu, D. Pasetto, C. Scudeler, C. Paniconi, M. Putti, P. A. Troch, S. B. DeLong, K. Dontsova, L. Pangle, D. D. Breshears, J. Chorover,
481 T. E. Huxman, J. Pelletier, S. R. Saleska, and X. Zeng. Incipient subsurface heterogeneity and its effect on overland flow generation - Insight
482 from a modeling study of the first experiment at the Biosphere 2 Landscape Evolution Observatory. *Hydrol. Earth Syst. Sci.*, 18:1873–1883,
483 2014. doi: 10.5194/hess-18-1873-2014.

- 484 [11] G. Evensen. *Data Assimilation: The Ensemble Kalman Filter*. Springer, New York, 2007.
- 485 [12] G. Evensen. The ensemble Kalman filter for combined state and parameter estimation. *IEEE CONTR. Syst. Mag.*, 29(3):83–104, 2009. doi:
486 10.1109/MCS.2009.932223.
- 487 [13] M. Camporese, G. Cassiani, R. Deiana, and P. Salandin. Assessment of local hydraulic properties from electrical resistivity tomography
488 monitoring of a three-dimensional synthetic tracer test experiment. *Water Resour. Res.*, 47(12):W12508, 2011. doi: 10.1029/2011WR010528.
- 489 [14] E. Crestani, M. Camporese, D. Baú, and P. Salandin. Ensemble Kalman filter versus ensemble smoother for assessing hydraulic conductivity
490 via tracer test data assimilation. *Hydrol. Earth Syst. Sci.*, 17(4):1517–1531, 2013. doi: 10.5194/hess-17-1517-2013.
- 491 [15] H.-J. Hendricks Franssen and W. Kinzelbach. Ensemble Kalman filtering versus sequential self-calibration for inverse modeling of dynamic
492 groundwater flow systems. *J. Hydrol.*, 365:261–274, 2009. doi: 10.1016/j.jhydrol.2008.11.033.
- 493 [16] J. Carrera and S. P. Neumann. Estimation of aquifer parameters under transient and steady state conditions: 2. Uniqueness, stability and
494 solution algorithms. *Water Resour. Res.*, 22(2):211–227, 1986.
- 495 [17] J. P. Drecourt, H. Madsen, and D. Rosbjerg. Calibration framework for a Kalman filter applied to a groundwater model. *Adv. Water Resources*,
496 29(5):719–734, 2006. doi: 10.1016/j.advwatres.2005.07.007.
- 497 [18] G. J. M. De Lannoy, R. H. Reichle, P. R. Houser, V. R. N. Pauwels, and N. E. C. Verhoest. Correcting for forecast bias in soil moisture
498 assimilation with the ensemble Kalman filter. *Water Resour. Res.*, 43(9):W09410, 2007.
- 499 [19] H.-J. Hendricks Franssen and W. Kinzelbach. Real-time groundwater flow modeling with the ensemble Kalman filter: Joint estimation of
500 states and parameters and the filter inbreeding problem. *Water Resour. Res.*, 44:W09408, 2008. doi: 10.1029/2007WR006505.
- 501 [20] Y. Chen and D. S. Oliver. Cross-covariances and localization for EnKF in multiphase flow data assimilation. *Comput. Geosci.*, 14:579–601,
502 2010. doi: 10.1007/s10596-009-9174-6.
- 503 [21] J. L. Anderson. Localization and sampling error correction in ensemble Kalman filter data assimilation. *Mon. Weather Rev.*, 140(7):2359–
504 2371, 2012. doi: 10.1175/MWR-D-11-00013.1.
- 505 [22] A. Y. Sun, A. Morris, and S. Mohanty. Comparison of deterministic ensemble Kalman filters for assimilating hydrogeological data. *Adv.*
506 *Water Resources*, 32(2):280 – 292, 2009. doi: 10.1016/j.advwatres.2008.11.006.
- 507 [23] A. Y. Sun, A. Morris, and S. Mohanty. Sequential updating of multimodal hydrogeologic parameter fields using localization and clustering
508 techniques. *Water Resour. Res.*, 45:W07424, 2009. doi: 10.1029/2008WR007443.
- 509 [24] M. Panzeri, M. Riva, A. Guadagnini, and S. P. Neuman. Data assimilation and parameter estimation via ensemble Kalman filter coupled with
510 stochastic moment equations of transient groundwater flow. *Water Resour. Res.*, 49(3):1334–1344, 2013. doi: 10.1002/wrcr.20113.
- 511 [25] A. H. Alzraiee, D. Baú, and Elhaddad A. Estimation of heterogeneous aquifer parameters using centralized and decentralized fusion of
512 hydraulic tomography data. *Hydrol. Earth Syst. Sci.*, 18:3207–3223, 2014. doi: 10.5194/hess-18-3207-2014.
- 513 [26] Y. Chen and D. Zhang. Data assimilation for transient flow in geologic formations via ensemble Kalman filter. *Adv. Water Resources*, 29(8):
514 1107–1122, 2006. doi: 10.1016/j.advwatres.2005.09.007.
- 515 [27] M. Camporese, C. Paniconi, M. Putti, and P. Salandin. Ensemble Kalman filter data assimilation for a process-based catchment scale model
516 of surface and subsurface flow. *Water Resour. Res.*, 45:W10421, 2009. doi: 10.1029/2008WR007031.

- 517 [28] M. T. van Genuchten. A closed-form equation for predicting the hydraulic conductivity of unsaturated soils. *Soil Sci. Soc. Amer. J.*, 44:
518 892–898, 1980.
- 519 [29] D. Pasetto, M. Camporese, and M. Putti. Ensemble Kalman filter versus particle filter for a physically-based coupled surface-subsurface
520 model. *Adv. Water Resources*, 47(1):1–13, 2012. doi: 10.1016/j.advwatres.2012.06.009.
- 521 [30] H. Moradkhani, K.-L. Hsu, H. Gupta, and S. Sorooshian. Uncertainty assessment of hydrologic model states and parameters: Sequential data
522 assimilation using the particle filter. *Water Resour. Res.*, 41(5):W05012, 2005. doi: 10.1029/2004WR003604.
- 523 [31] A. H. Jazwinski. *Stochastic Processes and Filtering Theory*. Academic Press, New York, 1970.
- 524 [32] R. E. Kalman. A new approach to linear filtering and prediction problems. *J. Basic Eng.-T. ASME*, 82(Series D):35–45, 1960.
- 525 [33] G. Evensen. The ensemble Kalman filter: theoretical formulation and practical implementation. *Ocean Dyn.*, 53(4):343–367, 2003. doi:
526 10.1007/s10236-003-0036-9.
- 527 [34] A. I. Gevaert, A. J. Teuling, R. Uijlenhoet, S. B. DeLong, T. E. Huxman, L. Pangle, D. D. Breshears, J. Chorover, J. D. Pelletier, S. R. Saleska,
528 X. Zeng, and P. A. Troch. Hillslope-scale experiment demonstrates role of convergence during two-step saturation. *Hydrol. Earth Syst. Sci.*,
529 18:3681–3692, 2014. doi: 10.5194/hess-18-3681-2014.
- 530 [35] A. Bellin and Y. Rubin. HYDRO_GEN: A spatially distributed random field generator for correlated properties. *Stoch. Hydrol. Hydraul.*, 10
531 (4):253–278, 1996. doi: 10.1007/BF01581869.
- 532 [36] M. A. Cushy, A. Bellin, and Y. Rubin. Generation of three-dimensional flow fields for statistically anisotropic heterogeneous porous-media.
533 *Stoch. Hydrol. Hydraul.*, 9(1):89–104, 1995. doi: 10.1007/BF01581760.
- 534 [37] Y. Shi, K. J. Davis, F. Zhang, C. J. Duffy, and X. Yu. Parameter estimation of a physically based land surface hydrologic model using the
535 ensemble Kalman filter : A synthetic experiment. *Water Resour. Res.*, 50:706–724, 2014. doi: 10.1002/2013WR014070.